LETTER • OPEN ACCESS

Effects of rising atmospheric CO$_2$, climate change, and nitrogen deposition on aboveground net primary production in a temperate forest

To cite this article: Wen J Wang et al 2019 Environ. Res. Lett. 14 104005

View the article online for updates and enhancements.
Effects of rising atmospheric CO₂, climate change, and nitrogen deposition on aboveground net primary production in a temperate forest

Wen J Wang 1, 2, 3, 7, 8, Shuang Ma 1, 2, 7, Hong S He 1, 2, 8, Zhihua Liu 1, Frank R Thompson III 1, Wenchi Jin 1, Zheng Fang Wu 1, 2, Martin A Spetch 1, Lei Wang 1, Song Xue 1, Wenguang Zhang 1 and Xianwei Wang 1

1 Key Laboratory of Geographical Processes and Ecological Security in Changbai Mountains, Ministry of Education, School of Geographical Sciences, Northeast Normal University, Changchun, 130024, People’s Republic of China
2 Northeast Institute of Geography and Agroecology, Chinese Academy of Sciences, Changchun, 130102, People’s Republic of China
3 School of Natural Resources, University of Missouri, 203 ABNR Bldg, Columbia, MO, 65211, United States of America
4 CAS Key Laboratory of Forest Ecology and Management, Institute of Applied Ecology, Chinese Academy of Sciences, Shenyang, 110016, People’s Republic of China
5 USDA Forest Service, Northern Research Station, 202 ABNR Bldg, Columbia, MO, 65211, United States of America
6 Arkansas Forestry Science Laboratory, USDA Forest Service, Southern Research Station, PO Box 1270, Hot Springs, AR 71902 United States of America
7 Joint first authors: Wen J Wang and Shuang Ma contributed equally to this work.
8 Authors to whom any correspondence should be addressed.

E-mail: wangwenj@missouri.edu, heh@missouri.edu and wuzf@nenu.edu.cn

Keywords: carbon sink, forest productivity, ecosystem modeling, PnET, net primary production (NPP), temperate forest

Abstract

Temperate forests regrowing from historical land use and land cover change in the eastern US serve as carbon (C) sinks. Environmental drivers have been significantly altered (e.g. rising atmospheric CO₂ concentration, warmer temperature, and elevated nitrogen (N) deposition) and will have a wide range of impacts on future forest C sinks. However, the interactions among these environmental drivers are unclear and their effects are subject to uncertainty. We assessed the combined and interactive effects of rising CO₂ concentration, climate change (temperature, precipitation), and N deposition on forest aboveground net primary production (ANPP) and their relative contribution to ANPP changes of a temperate forest in the eastern US. We used a process-based ecosystem model PnET-day to simulate coupled cycles of C, water, and N of forest ecosystems. We found that (1) climate change exerted negative effects on ANPP (−0.250 kg C m⁻² yr⁻¹) whereas rising CO₂ and N deposition enhanced ANPP (+0.014 kg C m⁻² yr⁻¹); (2) climate change interacted with rising CO₂ and N deposition to decrease ANPP (−0.032, −0.018 kg C m⁻² yr⁻¹); rising CO₂ and N deposition acted in synergy to increase ANPP (+0.014 kg C m⁻² yr⁻¹); (3) changes in ANPP were mainly attributed to rising CO₂ and climate change whereas N deposition effects and any two- or three-factor interactive effects were relatively small. Our results suggest that the total negative effect sizes will not be offset by the total positive effect sizes, thus resulting in reductions in forest ANPP over the 21st century.

1. Introduction

Temperate forests, making up 25% of all forest area, sequestrate approximately 30% global carbon (C) emissions and store about 15% of global C (Martin et al 2001). The C sinks taken up by temperate forests have increased in the past decades primarily because of increasing density of biomass and great increases in forest area (Pan et al 2011). Recent studies generally agree that contemporary C sinks of temperate forests in eastern US have been largely attributed to forest regrowth from historical land use and land cover change such as forest clearing, cropland expansion, and followed agricultural abandonment (Caspersen et al 2000, Pacala et al 2007). However, the future persistence and magnitude of C sinks of the temperate forests in the eastern US remain uncertain.

Environmental conditions have been significantly altered by human-induced activities over the past century and will continue to change in the future such as...
rising atmospheric CO2 concentration, warmer temperature, altered precipitation regimes, and elevated nitrogen (N) deposition (IPCC 2014). Both individually and interactively, these changes of environmental drivers have a wide range of impacts on future forest C sinks (Fekete et al 2017, Houtven et al 2019). Rising CO2 concentration and warmer temperature will favor net primary productivity (NPP) and water use efficiency of trees and thus promote C sequestration (Fang et al 2014, Fernández-Martínez et al 2019), though the extent of CO2 fertilization effects will depend upon other factors such as nutrients (de Vries et al 2017). Altered precipitation regimes such as increased drought stress will decrease tree growth and even trigger tree mortality under severe water limited conditions (Ibáñez et al 2018). Elevated N deposition is almost likely to increase tree growth, thus leading to higher C accumulation given that N is often the limiting nutrient in many forest ecosystems (Feng et al 2015, Maes et al 2019). However, the fertilization effects of elevated N deposition on tree growth are expected to saturate or even likely to decline with continuous N enrichment because of nonlinear responses of tree growth to accumulated N deposition (Tian et al 2016). These environmental drivers are also believed to act in synergy resulting significant interactive effects (Norby et al 2010, Dusenge et al 2019). For example, concurrent rise in CO2 concentration and N deposition could bring out strong synergistic fertilization effects especially in the forest ecosystems with N limitation (O’Sullivan et al 2019). Understanding the combined effects of these environmental drivers is, therefore, needed to generate reliable projections of future forest C dynamics, which are essential for land managers and policy makers to design effective plans for natural resource conservation and adaptation strategies for climate change.

In our study, we assessed the combined effects of rising atmospheric CO2 concentration, climate change (temperature, precipitation), and N deposition on aboveground net primary production (ANPP) of a temperate forest in the eastern US. We used a process-based ecosystem model PnET-day to simulate physiological responses of forest ecosystems to rising atmospheric CO2, climate change, and elevating N deposition. We used PnET-day model because it simulates the coupled cycles of C, water, and N of forest ecosystems driven by daily environmental data (Aber et al 1996). It has been validated against inventory data and eddy-flux tower measurements and applied to various forests (Aber and Federer 1992, Aber et al 1996, Ryu et al 2008, Jin et al 2016). We addressed the following questions: (1) how forest ANPP will change under rising CO2, climate change, and N deposition over the 21st century? and (2) what is the relative contribution of rising CO2, climate change, N deposition, and the interaction of any two- and three-combination of these three environmental drivers on forest ANPP changes?

2. Data and methods

2.1. Study area

Our study area included the Current River Hills ecological subsection (223 Af) within the Ozark Highland ecological section (223A) (McNab et al 2007) and was located in the US state of Missouri (figure 1). The study area covered approximately 810 000 hectares. It is a mature dissected plateau with dolomite and sandstone bedrock, and soils primarily developed from cherty limestones (McNab et al 2007). The climate is continental temperate with warm and humid summers and relatively dry and cold winters. The average annual temperature is 13.4 °C and average annual precipitation is 1100 mm.

We chose the Current River Hills ecological subsection as the study area because it contained one of the largest blocks of forest in the eastern US (figure 1). This area was primarily composed of secondary upland oak-hickory and oak-pine forests. The dominant tree species included white oak (Quercus alba L.), black oak (Q. velutina Lamb.), scarlet oak (Q. coccinea Muenchh.), bitternut hickory (Carya cordiformis (Wangenh.) K. Koch), and pignut hickory (C. glabra Miller) (Shifley and Brookshire 2000). Oaks and hickories have been the dominant tree species for the past thousands of years and are the keystone species of major importance in maintaining biodiversity, ecosystem functions and services. Similar to many forests in the eastern US, current forests in the study area are regrowing from previous disturbances (mainly tree harvesting and agricultural abandonment) and the average stand age is approximately 70 years.

2.2. Modeling approach and experimental design

We designed a three-factor factorial experimental design with two levels of CO2 concentration (current CO2 concentration and future rising CO2 concentration), four levels of climate (current climate and three climate change scenarios), and two levels of N deposition (current N and future elevated N deposition), resulting in total of 16 simulation scenarios. We treated the current CO2 concentration, climate (temperature and precipitation), and N deposition scenario as the baseline scenario. We used PnET-day model to simulate the C, N and water dynamics of forest ecosystems for the 16 simulation scenarios.

2.2.1. Climate data and climate change scenarios

We included current climate and three climate change scenarios based on CanESM2, GFDL-ESM2M, and MIROC5 global climate models (GCMs) under the representative concentration pathway (RCP) 8.5 (Riahi et al 2007). Since the three GCMs projected different future climate change patterns, we were able
Figure 1. The study area included the Current River Hills ecological subsection (223 Af) within the Ozark Highland ecological section (223A) and was located in the US state of Missouri and dominated by oak-hickory forests.

Figure 2. Annual maximum daily temperature and annual precipitation projected by three global climate models (CanESM2, GFDL-ESM2M, and MIROC5) and their mean values under RCP 8.5, future rising CO₂ concentration, and elevated nitrogen deposition under RCP 8.5 from 2010 to 2099 for the Current River Hills ecological subsection.
to generate a range of projections to incorporate uncertainties in future climate projections (figure 2, Rupp et al 2013).

We used daily maximum and minimum temperature, precipitation, and solar radiation for each climate scenario. We obtained the current climate data (1980–2009) for our study area at 1 km resolution from DAYMET (Thornton et al 2016). We downloaded the future climate data (2010–2099) for three climate change scenarios from the Coupled Model Intercomparison Project phase 5 (CMIP5, https://cmip.llnl.gov/cmip5/data_portal.html). Compared to the current climate scenario, three climate change scenarios projected the annual maximum daily temperature to increase by 4 °C–6 °C in 2070–2099 (figure 2), in which the CanESM2 and MIROC5 RCP 8.5 scenarios showed more dramatic increases than the GFDL-ESM2M RCP 8.5 scenario. Annual precipitation on average had relatively small anomalies from current climates with increased annual variations after 2050 (figure 2).

2.2.2. CO2 concentration scenarios
We obtained the annual CO2 concentration for the current CO2 scenario (1980–2009) from Mauna Loa Observatory (http://esrl.noaa.gov/gmd/obop/mlo/). We used the atmospheric CO2 concentration projection (2010–2099) under RCP 8.5 as the future rising CO2 scenario (https://cmip.llnl.gov/cmip5/data_portal.html). CO2 concentration under RCP 8.5 was projected to continue rising over 900 ppm and averaged 600 ppm over the 21st century (figure 2).

2.2.3. N deposition scenarios
We downloaded the N deposition data (1980–2009) for the current N deposition scenario from the National Atmospheric Deposition Program (http://nadp.sws.uiuc.edu/committees/tdep/tdepmaps/). Since RCP 8.5 did not have N deposition projections to the end of this century, we used the projections of atmosphere mass content of nitrate dry aerosol from 2006 to 2100 to derive the future N deposition based on the proportional relationship between N deposition and the atmosphere mass content of nitrate dry aerosol (Elliott et al 2009). Projected N deposition increased to 2040 followed by a decreasing trend and averaged 2.02 g m^-2 yr^-1 over the 21st century (figure 2).

2.3. PnET-day parameterization and validation
We used PnET-day model, an improved version of PnET-CN, to simulate the effects of rising CO2 concentration, climate change, and N deposition on forest ANPP (http://pnet.sr.unh.edu/, Aber et al 1996, Ollinger et al 2002). PnET family models are process-based ecosystem models that simulate the C, water, and N dynamics of forest ecosystems (Aber and Federer 1992, Ollinger et al 2002). The core of PnET-day was a daily time-step canopy submodel that simulated physiological processes of a multi-layered canopy using generalized leaf-trait relationships and vertical gradients in physical environmental conditions (Aber et al 1996). PnET-day model represented the vegetation using three major components: leaves, wood, and roots. It simulated the maximum potential net photosynthetic rate under light-saturated conditions as a function of foliar N concentration and simulated water use efficiency (mg C fixed per g H2O transpired) as a function of vapor pressure deficit. Based on these two principles, a link between C dynamics and water transpiration was established and the computation load for water transpiration was greatly reduced because it only depended on vapor pressure deficit. A complete N cycle in PnET-day model included N mineralization, nitrification, plant N uptake and leaching losses. Since PnET-day model simulated the interactions between C and N cycles, forest C sequestration was affected by both plant and soil N pools. The annual net primary productivity was summarized based on the net photosynthetic rate and the annual aboveground net primary productivity was allocated from the annual net primary productivity (Aber et al 1996).

We obtained the average water holding capacity for the entire subsection from Soil Survey Geographic Database (SSURGO, Soil Survey Staff 2014). We initialized the forests using average C mass of leaves, wood, and roots from forest inventory and analysis (FIA) data (O’Connell et al 2015). We used the generalized parameters for the broad-leaved deciduous forests (e.g. light attenuation constant, Q10 for leaf respiration, minimum and optimum daytime temperature) from previous studies (Reich et al 1995, Aber et al 1996, Ollinger et al 2002). We derived the site-specific parameters (e.g. foliar N content, specific leaf weight, growing degree days to start or complete foliar production, max. or min. winter foliar biomass) based on previous studies (Aber and Federer 1992, Tjoelker et al 2001, Scheller and Mladenoff 2005, Ryu et al 2008, Xu et al 2009) and the eddy-flux tower measures in the Missouri Ozarks AmeriFlux (MOFLUX) site (doi.org/10.17190/AMF/1246081).

We used historical FIA data to validate mode simulation results. We firstly initialized the forest ecosystems in PnET-day model using 136 FIA plots from the 2000 survey cycle and then run the model from 2000 to 2010 under the baseline scenario (with current climate, CO2 and N deposition). We compared the average aboveground ANPP predicted by PnET-day model with the average ANPP estimated from FIA data over a 10 year period (2000–2010). Since FIA data did not contain ANPP information, we calculated three components: net C density increment from 2000 to 2010, aboveground mortality, and aboveground C removed due to disturbances; we then divided the sum of these three components by a time interval of 10 years to obtain the average ANPP. Predicted
average ANPP from PnET-day from 2000 to 2010 was 0.551 kg C m$^{-2}$ yr$^{-1}$, which was comparable to the observed average ANPP (0.576 kg C m$^{-2}$ yr$^{-1}$) from FIA data for the same period. Thus, PnET-day model simulated consistent ANPP with observed field inventories in our study area. Finally, we reinitialized the PnET-day model using 142 FIA plots in the 2010 survey cycle and then simulated forest ecosystems from 2010 to 2099 for the 16 simulation scenarios.

2.4. Analysis of simulation results
We calculated the $\Delta$ANPPs for the simulation scenario $i$ as the difference in ANPP between the simulation scenario $i$ and the baseline scenario (ANPP$_{i}$ − ANPP$_{\text{baseline}}$). We also calculated the average $\Delta$ANPPs for the period 2010–2100 for each simulation scenario. We averaged the $\Delta$ANPPs for the three climate change scenarios to focus on an average response, which was referred to as mean RCP 8.5 results.

We estimated the single-factor effect size of rising CO$_2$ concentration, climate change, or N deposition as the difference in $\Delta$ANPPs from 2010 to 2009 between the baseline scenario and the simulation scenario in which the single factor was only included; We calculated any two- and three-factor combined effect as the difference in $\Delta$ANPPs between the baseline scenario and the simulation scenario in which the given two or three factors were included simultaneously; We calculated any two-factor interactive effect by subtracting both single-factor effects from the given two-factor combined effect; we finally calculated the three-factor interactive effect by subtracting both the three single-factor effects and three two-factor interactive effects from the three-factor combined effect (Temperli et al. 2013). We averaged the single-factor effects, two-and three-factor interactive effects for the period 2010–2100 to represent the average effect sizes. Note that our analyses focused on average responses among three climate change scenarios and effect sizes rather than statistical significance.

3. Results

3.1. Temporal changes in ANPP under future scenarios
$\Delta$ANPPs varied greatly annually (figure 3) and averaged $-0.041$ kg C m$^{-2}$ yr$^{-1}$ under the RCP 8.5 climates, future rising CO$_2$ and N deposition scenario over the 21st century (figure 4). $\Delta$ANPPs on average were overall less with climate change and greater with rising CO$_2$ and N deposition (figure 3). With climate change, the average $\Delta$ANPP was reduced by 0.250 kg C m$^{-2}$ yr$^{-1}$ under the mean RCP8.5 climate scenario followed by GFDL-ESM2M and MIROC5 RCP 8.5 climate scenarios, in comparison with the baseline scenario (figures 3 and 4). Future rising CO$_2$ and N deposition increased the $\Delta$ANPP by 0.253 and 0.014 kg C m$^{-2}$ yr$^{-1}$, respectively, compared with the baseline (figures 3 and 4). Climate change interacted with rising CO$_2$ and N deposition to reduce the average $\Delta$ANPP by 0.029 and 0.255 kg C m$^{-2}$ yr$^{-1}$, respectively, while rising CO$_2$ interacted with N deposition to increase the average $\Delta$ANPP by 0.28 kg C m$^{-2}$ yr$^{-1}$ (figure 4). The interaction of climate change, rising CO$_2$ and N deposition resulted in reductions in average $\Delta$ANPP by 0.041 kg C m$^{-2}$ yr$^{-1}$ (figure 4).

3.2. Interactive effects of rising CO$_2$, climate change, and N deposition
The effect sizes of climate change on $\Delta$ANPP overall increased with great annual variation and averaged $-0.250$ kg C m$^{-2}$ yr$^{-1}$ whereas the effect sizes of rising CO$_2$ continued increasing and averaged 0.253 kg C m$^{-2}$ yr$^{-1}$ (figures 5 and 6). The effect sizes of N deposition on $\Delta$ANPP increased by mid-century followed a decreasing trend and averaged 0.014 kg C m$^{-2}$ yr$^{-1}$ (figures 5 and 6). The average effect sizes for the interaction of rising CO$_2$ and climate change and the interaction of climate change and N deposition on $\Delta$ANPP were $-0.032$ and $-0.018$ kg C m$^{-2}$ yr$^{-1}$, respectively (figure 6). The effect sizes for the interaction of rising CO$_2$ and N deposition on $\Delta$ANPP averaged 0.014 kg C m$^{-2}$ yr$^{-1}$ (figure 6). The average effect sizes for the interaction of climate change, rising CO$_2$, and N deposition on $\Delta$ANPP were $-0.020$ kg C m$^{-2}$ yr$^{-1}$ (figure 6). Thus, on average, the changes in ANPP were mainly attributed to rising CO$_2$ and climate change while the interactive effects of any two- and three-combination of rising CO$_2$, climate change, and N deposition were relatively small. The total positive effect sizes (e.g. rising CO$_2$, N deposition, the interaction of rising CO$_2$ and N deposition) were not enough to offset the total negative effect sizes (e.g. climate change, interaction of rising CO$_2$ and climate change), thus resulting in reductions in ANPP over the 21st century.

4. Discussion
It is important to note that our simulations are not to be interpreted as forecasts of future. Rather, we assessed the combined effects of three environmental drivers (rising CO$_2$ concentration, climate change, and N deposition) on forest ANPP and their relative contribution to ANPP changes. We found rising CO$_2$ and N deposition enhanced ANPP whereas climate change exerted negative effects on ANPP; rising CO$_2$ and N deposition acted in synergy to increase ANPP; climate change interacted with rising CO$_2$ or/and N deposition to decrease ANPP; future changes in ANPP were mainly attributed to rising CO$_2$ and climate
change whereas N deposition effects and any two- or three-factor interactive effects were relatively small.

The largest positive effects on ANPP was found for rising CO2 concentration among three environmental drivers, which agree with many studies (Fernández-Martínez et al 2019, Tharammal et al 2019). Our results showed that, under current climates and N deposition level, ANPP with elevated CO2 concentrations of 600 ppm increased by 0.253 kg C m$^{-2}$ yr$^{-1}$ (21%). These estimates were similar to those of free-air CO2 enrichment (FACE) experiments. For example, the FACE experiments conducted in four temperate forest sites in the Northern Hemisphere showed an average increase of 18% in NPP under the elevated CO2 concentration (approx. 550 ppm) (Norby et al 2010). However, it is difficult to compare our results under higher CO2 concentrations (>550 ppm) in a long timeframe with field-based forest ecosystem studies. We also found although CO2 concentration kept increasing after 2080, most of CO2 fertilization effects were found in the period of 2040–2080 with much smaller changes in ΔANPP after 2080, suggesting CO2 saturation effects. Enhanced photosynthesis under elevated CO2 concentration may be reduced via physiological adjustments or limited by N availability (Voelker et al 2016). It should be noted that the PnET model, similar to many other process-based models, scales up the leaf level responses to CO2 fertilization directly to the ecosystem level, and thus our results should be considered to be at the upper bound of possible responses of forest ecosystems to rising CO2 (Ollinger et al 2008).

Climatic warming had a secondary effects on forest ANPP changes. ANPPs under the RCP 8.5 climates declined compared to those under current climates. Our results were consistent with predictions by Peters et al (2013) using PnET-CN model in Great Lakes forests in the northern US under the GFDL A1FI climate change scenario over the 21st century. However, Peters et al (2013) showed that ANPP was projected to slightly increase under the moderate climate change scenario PCM B1, which might indicate the saturation effects of warming climates. Moderate warming climates could enhance forest productivity by accelerating photosynthesis (Fang et al 2014, Bowling et al 2018) and lengthening growing season (Babst et al 2019). However, under more severe climate scenarios (e.g. RCP 8.5, GFDL A1FI), the negative effects of climate change could be because the elevated temperature was beyond the optimal temperature of photosynthesis.

Figure 3. ΔANPP (kg C m$^{-2}$ yr$^{-1}$) predicted by PnET-day model under alternative simulation scenarios: rising CO2 concentration alone, N deposition alone, both rising CO2 and N deposition, climate change alone, both climate change and rising CO2, both climate change and N deposition, and all climate change, rising CO2 and N deposition from 2010 to 2099 for the Current River Hills ecological subsection, compared to the baseline scenario.
and thus caused photosynthetic rate to decline (Bagley et al. 2015), and/or that the respiration increased with temperature (Ballantyne et al. 2017), and/or that water was limited under extreme climate change scenario (Helman et al. 2017), among other factors.

Among rising CO₂ concentration, climate change, and N deposition, our study suggested that N deposition had the smallest effect on ANPP (+0.043 kg C m⁻² yr⁻¹). The interactive effects of N deposition with climate change and/or CO₂ fertilization were also relatively small compared to effects of CO₂ fertilization, climate change, and interaction of climate change and CO₂ fertilization. Our results generally agreed with some previous studies showing that
N deposition made only minor contributions to C sequestration in temperate forests (e.g. Fernández-Martínez et al. 2017). However, our results differed with some field experimental studies of N fertilization (e.g. Lovett et al. 2013, Etzold et al. 2014) because most experiments had higher N deposition rates than those used in our study. For example, Lovett et al. (2013) conducted 6 year increased N deposition experiments in southeastern New York State with an increased N deposition rate of 5 g N m\(^{-2}\) yr\(^{-1}\), which did not significantly enhance ANPP for any tree species but significantly increased C stock in forest soils. This may be because only a small fraction of added N entered the trees, while most of added N remained in soil organic matter (Goodale 2017). This finding was supported by Etzold et al. (2014) which showed that N deposition enhanced forest productivy with an increased N deposition rate of 2–3 g N m\(^{-2}\) yr\(^{-1}\) and no further growth increased beyond that threshold. N can be absorbed directly into trees through canopies in temperate forests (Guerrieri et al. 2015, Wang et al. 2017a), however, in most forest models (including PnET) and field experiments, N enters the forest ecosystem in the forest soil surface, the effect of which may be different from that of equivalent N addition from atmospheric deposition.

Our study suggest the negative interactive effects of rising CO\(_2\) and climate change on forest ANPP, which were consistent for observational and modeling studies that indicated the CO\(_2\) fertilization effects on forest production were constrained by climate warming (e.g. Fernández-Martínez et al. 2019). The negative interactive effects may arise because of higher respiratory rates and drought stress under future warmer and drier climates. Rising CO\(_2\) were expected to interact with elevated N deposition to increase forest production because N addition moderated N limitation (Norby et al. 2010). Many forest ecosystem models are considered to have a tendency to overestimate the CO\(_2\) fertilization effects because N limitation is not included in model formulations (He et al. 2017, Faticchi et al. 2019). However, PnET-day model was able to capture the effects of N on photosynthetic rate through simulating N cycles and its interaction with C cycles (Ollinger et al. 2002).

It should be noted that there are several limitations in this study. First, PnET treats forest ecosystem as a lumped system consisting of different plant organs (e.g. roots, leaves), neither individual plant species nor functional types were represented; thus forest demographic processes and composition changes cannot be simulated. Most of temperate forests are relatively young and are typically ‘regrown’ from historical land management activities (e.g. harvest, agricultural abandonment) and are expected to continue taking up large amounts of C in the future (Reich and Frelich 2002). Recent studies suggest that age-driven forest demographic processes alone might play a larger role than growth enhancement from environmental changes (e.g. rising CO\(_2\)) (Liang et al. 2015, Jin et al. 2017, Wang et al. 2017b, 2018, 2019, Pugh et al. 2019). Secondly, PnET does not simulate effect of nutrients other than N, e.g. phosphorus (P), on ANPP. It has been suggested the models neglecting P limitations.
may overestimate terrestrial C uptake (Fernández-Martínez et al. 2017) and there could be interactions between N and P cycles (Schulte-Uebbing and de Vries 2017). Nevertheless, our modeling approach is well suited for assessing the effects of rising atmospheric CO$_2$ concentration, climate change, and N deposition on forest production. PnET-day model can be directly parameterized from forest inventory data and has been well validated against inventory data and eddy-flux tower measurements (e.g. Ryu et al. 2008). With our modeling approach, we were able to incorporate the physiological responses of forest ecosystems to environmental changes through simulating coupled cycles of C, water, and N that would not be possible with some other models.

Acknowledgments

Funding for this work was provided by the National Natural Science Foundation of China (No. 41871045, No. 41801081), the Chinese Academy of Sciences (Y17H0301001), the Natural Science Foundation of Julian Province (201902011191C, 20190103153JH), Key laboratory project of wetland ecology and environment, Chinese Academy of Sciences (O9B3017001), the USDA Forest Service Northern Research Station, the Department of Interior Northeast Climate Science Center graduate fellowship, and the University of Missouri-Columbia. Any data that support the findings of this study are included within the article.

References

Aber J D and Federer C A 1992 A generalized, lumped-parameter model of photosynthesis, evapotranspiration and net primary production in temperate and boreal forest ecosystems Oecologia 92 463–74

Aber J D, Reich P B and Goulden M L 1996 Extrapolating leaf CO$_2$ exchange to the canopy: a generalized model of forest photosynthesis compared with measurements by eddy correlation Oecologia 106 252–65


Ballantyne A et al. 2017 Accelerating net terrestrial carbon uptake during the warming hiatus due to reduced respiration Nat. Clim. Change 7 748–52


Dusenge M E, Duarte A G and Way D A 2019 Plant carbon metabolism and climate change: elevated CO$_2$ and temperature impacts on photosynthesis, photosorption and respiration New Phytol. 221 32–49


Fernández-Martínez M et al. 2017 Atmospheric deposition, CO$_2$, and change in the land carbon sink Sci. Rep. 7 9632


Goodale C L 2017 Multiyear fate of a N-15 tracer in a mixed deciduous forest: retention, redistribution, and differences by mycorrhizal association Glob. Change Biol. 23 867–80


Helman D, Oseyn Y, Yakir D and Lensky M I 2017 Relationships between climate, topography, water use and productivity in two key Mediterranean forest types with different water-use strategies Agric. For. Meteorol. 232 319–30

Houtven G V, Phelan J, Clark C, Sabo R D, Buckley J, Thomas R Q, Horn K and Ledu S D 2019 Nitrogen deposition and climate change effects on tree species composition and ecosystem services for a forest cohort Ecol. Monogr. 0 e01345


Jin W C, He H S and Thompson F R 2016 Are more complex physiological models of forest ecosystems better choices for plot and regional predictions? Environ. Modelling Softw. 75 1–14


O’Connell B M 2015 The Forest Inventory and Analysis Database: Database Description and User Guide for Phase 2, version 6.0.1. Forest Inventory and Analysis National Program (www.fia.fs.fed.us/library/database_documentation/)
O’Connell B M 2015 The Forest Inventory and Analysis Database: Database Description and User Guide for Phase 2, version 6.0.1. Forest Inventory and Analysis National Program (www.fia.fs.fed.us/library/database_documentation/)
Pan Y et al 2011 A large and persistent carbon sink in the world’s forests Science 333 986–91
Pugh T et al 2019 Role of forest regrowth in global carbon sink dynamics Proc Natl Acad. Sci. 116 4382–7
Reich P B, Walters M B, Kloppe1 B D and Ellsworth D S 1995 Different photosynthesis–nitrogen relations in deciduous hardwood and evergreen coniferous tree species Oecologia 104 24–30
Scheller R M and Mladenoff D J 2005 A spatially interactive simulation of climate change, harvesting, wind, and tree species migration and projected changes to forest composition and biomass in northern Wisconsin, USA Glob. Change Biol. 11 307–21
Soil Survey Staff 2014 Keys to Soil Taxonomy 12th edn (Washington, DC: USDA-Natural Resources Conservation Service)
Thammarat T, Bala, G, Narayannappa D and Nemani R 2019 Potential roles of CO2 fertilization, nitrogen deposition, climate change, and land use and land cover change on the global terrestrial carbon uptake in the twenty-first century Clim. Dyn. 52 4393–408
Voelker S L et al 2016 A dynamic leaf gas-exchange strategy is conserved in woody plants under changing ambient CO2; evidence from carbon isotope discrimination in paleo and CO2 enrichment studies Glob. Change Biol. 22 889–902
Wang R et al 2017a Global forest carbon uptake due to nitrogen and phosphorus deposition from 1850 to 2100 Glob. Change Biol. 23 4854–72
Wang W J et al 2017b Changes in forest biomass and tree species distribution under climate change in the northeastern United States Landscape Ecol. 32 1399–413
Wang W J et al 2018 Population dynamics has greater effects than climate change on tree species distribution in a temperate forest region J. Biogeogr. 45 2766–78
Wang W J et al 2019 Climate change and tree harvest interact to affect future tree species distribution changes J. Ecol. 107 1901–17